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生物炭吸附水体中重金属机理与工艺研究进展

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摘要: 生物炭因其良好的表面特性和孔隙结构,广泛的原料来源和广阔的产业化发展前景,已成为当今环境、农业和能源等领域的研究热点。针对生物炭对水体重金属的吸附研究,本文基于生物炭原料和制备工艺的多样性,综合分析了国内外生物炭重金属吸附机理的研究成果,详细阐述、分析了5种吸附作用机制(物理吸附、静电作用、离子交换、络合反应和化学沉淀)及其相关表征手段;同时评述了吸附工艺条件和重金属种类对生物炭吸附重金属的影响;指出生物炭重金属吸附领域未来的研究中,应开展针对重金属吸附的生物炭原料特性及吸附产物的多维、微纳尺度表征方法研究。

关键词: 生物炭; 水体重金属; 吸附机理; 表征

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Review of Biochar as Adsorbent for Aqueous Heavy Metal Removal

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Abstract: Biochar has a bright prospect due to its good surface properties, pore structure and broad raw materials of its production. It has already been a hotspot in the fields of environment, agriculture and energy. It can be produced from a variety of biomass feedstock, such as agricultural residues, manures, wood, bone and so on, under oxygen-limited conditions. It has great potentials for heavy metal remediation and waste water treatment due to its unique properties, low price, easy processing and wide range of preparation material. The existing literature was incorporated to understand the overall sorption research of heavy metals on biochar adsorbents. However, there was still lack of information on the roles of different sorption mechanisms for biochar. The effects of feedstock materials and pyrolysis temperature on biochar characteristics and metal adsorption capacity were discussed. The interaction mechanisms between biochar and heavy metals, such as ion exchange, complexation, physical sorption, precipitation and electrostatic interactions were analyzed in detail. The influence of adsorption conditions and heavy metal species on heavy metals adsorbed by biochar was also included. At the end, the future research directions on sorption of heavy metals by biochar were proposed. The review would help to build important theory and methodology foundation for directly controlling production and scientific utilization of biochar.

Key words: biochar; heavy metal; adsorption mechanism; characterization

引言

重金属在环境中无处不在,危及人类健康^[1]。由于人类各种生产活动:冶炼加工、矿山开采、化工生产等,将含有重金属的废水排放到水环境中,造成了严重的水体重金属污染。重金属污染与其他有机化合物污染不同,许多有机化合物可以通过自然界

本身物理的、化学的或生物的方式净化,使有害性降低或解除。重金属具有持久性和富集性,很难在环境中降解,对人类及动植物危害极大^[2]。环境中的铅、铜含量超标会显著抑制动植物生长,严重时甚至导致其死亡^[3];同时,重金属还会通过食物链进入人体,干扰人体正常的生理功能进而对其健康造成危害^[4-5]。因此,水体中重金属污染物的治理问题

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目前,移除和回收水体中重金属的方法主要有化学沉淀、氧化还原、离子交换、生物过滤、活性污泥法和吸附法等^[6-11]。其中,吸附法具有净化效果好、操作简便、不会造成二次污染等优点,被视为一种有效且高效的处理方式^[12-13],被广泛应用于城市废水、工业污水中重金属的去除^[14]。而决定其吸附效果和成本的核心因素是吸附剂^[15-16]。因此,研制低廉高效的吸附剂是当前水体污染物吸附处理问题的研究热点。

生物炭(biochar)是生物质在一定温度和缺氧条件下热裂解形成的一类高度芳香难溶性固体物质^[17-18]。通常生物炭具有一定的比表面积和丰富的极性官能团,例如羧基、酚羟基和氨基,且孔隙结构发达^[19-21],其作为一种潜在的新型重金属生物吸附材料已成为科技工作者的研究热点^[22]。与目前水体污染物吸附常用的活性炭吸附剂^[14]相比,生物炭未经活化处理,成本低,且来源更为广泛。生物炭的吸附应用是一个多赢策略,在治理水体中污染物的同时消除了焚烧、腐烂处理对环境造成的破坏^[23],并且生物炭生产过程中产生的生物油(bio-oil)和混合气(syngas)有助于缓解能源短缺^[24]。

国内外研究表明生物炭可用于吸附水溶液中的Pb(II)、Cu(II)、Zn(II)、Ni(II)、U(VI)、Ag(II)、Cr(V)和Cd(II)等金属离子^[25-29],但由于生物质原料特性迥异^[30-32],炭化工艺与转化过程复杂^[33-35],从而导致生物炭特性多变,可控性较弱。同时,生物炭的特性^[36-37]和不同的吸附条件^[38-39]以及目标重金属种类^[40-41]均显著影响其吸附行为和效果。由此可见,生物炭吸附水中重金属的影响因素众多,并存在复杂的吸附机理。

1 生物炭原料种类、制备工艺及其特性

生物炭的来源非常广泛,主要包括农林业废弃物如秸秆、草本、果壳、果皮果渣等,工业和城市生活中产生的有机废弃物如生活垃圾、污泥等,养殖废弃物如畜禽粪便等。这些生物质都被热转换为各种理化特性的生物炭并应用于水体、土壤等环境修复和改良。不同种类生物质制备的生物炭在元素组成、工业组成、矿质元素、比表面积、孔容、孔径、灰分含量、羧酸酯化、芳香化、脂肪族链状结构等理化性质上存在差异。

制备工艺的不同也会直接导致生物炭特性的差异,生物炭制备工艺参数中影响最为显著的是炭化温度^[32,42-44]。随着炭化温度的升高,生物炭的C元

素含量升高,N元素含量无明显变化,H、O元素含量下降。因此,H/C、O/C、(O+N)/C等含量比值下降,这说明随着炭化温度的升高,炭化程度、芳香化程度增强,极性减弱^[37,45]。脂肪型C—C键逐渐消失,取而代之的是芳香性C=C键逐渐增强。除此之外,随着炭化温度的升高,pH值升高^[46-47];灰分比重增多,P、K、Ca、Mg比重增多^[48];比表面积和总孔隙体积增大^[49-50];孔隙宽度减小^[51];产率下降;含氧官能团含量下降;阳离子交换能力(CEC值)下降^[48]。

2 生物炭重金属吸附效果分析

生物炭原料和炭化温度显著影响其重金属吸附的能力,同一生物炭对不同重金属的吸附效果不同。研究显示,目前用于水体中重金属吸附的生物炭原料来源可分为秸秆类、畜禽粪便类、果壳类、果皮果渣类、木本类、草本类、泥类、水生植物类、骨类。用于吸附水体中重金属离子涉及Pb(II)、Cu(II)、Zn(II)、Cr(III)、Cr(VI)、Cd(II)、As(V)、Ni(II)、Hg(II)等^[52-56],其中对Pb(II)、Cu(II)、Zn(II)、Cd(II)的吸附量较大^[57-59]。对Pb(II)有较好吸附效果的是畜禽粪便类生物炭、秸秆类生物炭和骨类生物炭,例如400℃碳化的猪粪生物炭对Pb(II)最大吸附量为230.7 mg/g^[60](表1);对Cu(II)具有良好吸附效果的是畜禽粪便类生物炭、秸秆类生物炭和草本类生物炭,例如400℃炭化猪粪生物炭对Cu(II)最大吸附量为88.23 mg/g^[60];对Zn(II)具有良好吸附效果的是畜禽粪便类生物炭、果壳类生物炭、秸秆类生物炭和水生植物类生物炭,例如猪粪在400℃制备条件下对Zn(II)的最大吸附量为79.62 mg/g^[60];对Cd(II)具有良好吸附效果的是畜禽粪便类生物炭、秸秆类生物炭、水生植物类生物炭和骨类生物炭,例如猪粪在400℃制备条件下对Cd(II)的最大吸附量为117.01 mg/g^[60]。综上所述,畜禽粪便类和秸秆类生物炭是极具潜力的重金属生物吸附剂。但值得注意的是最大吸附量不仅与生物炭原料、制备条件相关,同时也受到吸附条件(初始pH值、温度、吸附剂浓度、吸附质浓度)等因素的影响。表1中的吸附量均是在特定制备和吸附条件下所得,可比性较差,后续应将生物质特性与其吸附效果进行相关性分析。

3 生物炭重金属吸附机制

由于生物炭结构组成复杂,其对水体中重金属的吸附机理也十分复杂。如图1所示,生物炭重金属吸附机制主要表现为:物理吸附、静电作用、离子

表 1 生物炭原料、炭化温度及重金属最大吸附量

Tab.1 Maximum adsorption of heavy metals in water by biochar with different raw materials and pyrolysis temperature

分类	制备原料	炭化温度/°C	吸附重金属	最大吸附量/(mg·g ⁻¹)	文献序号
作物秸秆等废弃物	水稻秸秆	700	Pb(II)	126.58	[61]
	玉米秸秆	700	Pb(II)	121.95	[61]
	小麦秸秆	350~500	Pb(II)	50.00	[62]
	花生秸秆	400	Cd(II)	89.60	[63]
	水稻秸秆	700	Cd(II)	60.61	[61]
	玉米秸秆	700	Cd(II)	39.68	[61]
	小麦秸秆	350~500	Cd(II)	5.95	[62]
	大豆秸秆	400	Cu(II)	53.12	[63]
	油菜秸秆	400	Cu(II)	37.76	[63]
	玉米秸秆	700	Zn(II)	23.03	[64]
	芦苇秸秆	500	Ni(II)	11.93	[65]
	木屑	700	Pb(II)	47.62	[61]
	松木	300	Pb(II)	4.25	[66]
	橡木皮	400~450	Cr(VI)	7.51	[67]
	橡木	400~450	Cr(VI)	4.93	[67]
	木屑	700	Cd(II)	6.67	[61]
	松木	300	Cu(II)	4.46	[68]
	落叶松	500	Cu(II)	0.476	[69]
	橡木	500	Cu(II)	0.307	[69]
	互花米草	400	Cu(II)	48.49	[70]
	柳枝稷	500	Cu(II)	0.28	[69]
	芒草	500	Cd(II)	13.24	[71]
	花生壳	350~500	Pb(II)	45.45	[62]
	非洲牧豆壳	350	Pb(II)	45.30	[72]
	稻壳类	300	Pb(II)	2.40	[66]
非洲牧豆壳	350	Cd(II)	38.30	[72]	
花生壳	350~500	Cd(II)	6.29	[62]	
花生壳	700	Zn(II)	24.90	[64]	
杏仁壳	650	Ni(II)	20.00	[73]	
畜禽粪便等废弃物	猪粪	400	Pb(II)	230.70	[60]
	牛粪	200	Pb(II)	132.80	[74]
	猪粪	400	Cd(II)	117.01	[60]
	牛粪	350	Cd(II)	51.40	[75]
	猪粪	400	Cu(II)	88.23	[60]
	牛粪	350	Cu(II)	54.40	[75]
	猪粪	400	Zn(II)	79.62	[60]
	牛粪	350	Zn(II)	32.80	[75]
	鸡粪	600	Ni(II)	10.94	[76]
	猪残体	600	Pb(II)	106.40	[77]
果皮果渣	橘子皮	500	Pb(II)	86.96	[79]
	菠萝皮	750	Cr(VI)	7.44	[51]
	柚子皮	400	Cd(II)	9.35	[80]
	消解甘蔗渣	600	Pb(II)	135.40	[81]
	甜菜渣	300	Cr(VI)	123.00	[82]
	水生植物	紫根风眼莲	500	Pb(II)	34.41
紫根风眼莲		500	Zn(II)	41.23	[83]
紫根风眼莲		500	Cu(II)	44.17	[83]
紫根风眼莲		500	Cd(II)	39.81	[84]

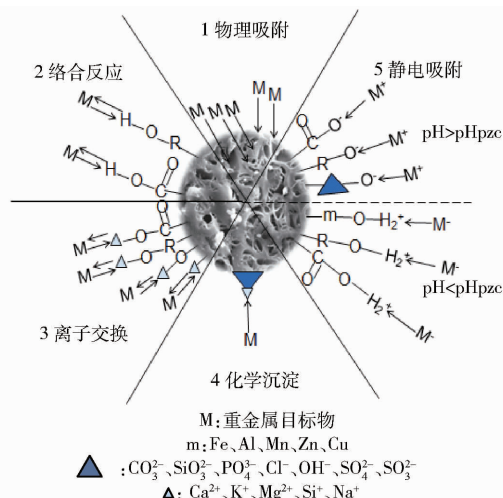


图 1 生物炭对水体中重金属的吸附机理示意图

Fig.1 Conceptual illustration of heavy metal sorption mechanisms on biochar

交换、络合反应和化学沉淀^[52-53,85]。生物炭的理化性质差异,吸附条件和重金属离子类型的变化都会造成吸附机理的不同,具体表现为每种机理对吸附过程的贡献不同。

3.1 物理吸附

物理吸附是由吸附质与吸附剂之间范德华力产生的,结合力较弱,属于可逆过程。生物炭粒子有着各种不同的比表面积和微孔,当制备温度在 500~700°C 时,更容易形成较高的比表面积和孔隙体积^[86]。这些结构可以与重金属离子发生物理吸附^[36,53,87-88]。

SHEN 等^[89]研究表明英国当地的硬木生物炭的粒径显著影响 Pb(II) 的吸附,0.15 mm 和 2 mm 生物炭对 Pb(II) 的最大吸附量分别为 47.66 mg/g 和 30.04 mg/g。生物炭在一定范围内,粒径减小,其比表面积会增加,进而提高对 Pb(II) 的吸附能力。但是比表面积大小并不是决定吸附强弱的重要理化参数^[10,61,90]。例如,畜禽粪便生物炭的比表面积随着炭化温度的升高而升高,但对重金属 Pb(II) 的吸附却未随比表面积的升高而增加。水稻秸秆生物炭比表面积为 73.40 m²/g,小于木屑生物炭的比表面积 391.12 m²/g,但对 Pb(II) 吸附量是水稻秸秆生物炭(126.58 mg/g)大于木屑生物炭(47.62 mg/g),对 Cd(II) 的吸附量也是水稻秸秆生物炭(60.61 mg/g)大于木屑生物炭(6.67 mg/g)。橡木生物炭的比表面积为 1~3 m²/g,其对 Cr(VI) 的移除效果与比表面积约为 1 000 m²/g 的活性炭相当^[10]。因此,生物炭表面的物理吸附占重金属吸附机理的比重较弱;同样的结果也被 SHEN 等^[89]和 ZHOU 等^[37]研究得到。

生物炭内部孔道也会发生物理吸附。LIU 等^[68]通过 2 种制备方式:300°C 水热炭化和 700°C 热

解分别制得2种松木生物炭,标记为H300和P700。测定了比表面积、平均孔直径和微孔体积差异。H300和P700的平均孔径分别为0.86、1.48 nm。戴静等^[61]分别测定了米糠、木屑、稻秆、玉米秸秆生物炭的平均孔径分别为2.15、2.03、3.56、2.33 nm。而金属离子Pb(II)、Cu(II)、Al(III)、Cr(III)、Cr(VI)、Ni(II)、Cd(II)、Zn(II)的直径分别为0.238、0.146、0.107、0.123、0.088、0.138、0.19、0.148 nm。这些重金属离子直径均小于生物炭的平均孔径。因此,在生物炭与重金属离子接触时,会有部分重金属离子进入到生物炭的孔隙中,不断填充孔隙,形成生物炭对重金属离子的物理吸附。一般来说,重金属的直径越小,渗透到生物炭孔径中越多,从而提高吸附量^[91-92]。

3.2 络合反应

除了物理吸附之外,其他化学吸附机制如表面络合反应也参与重金属离子的吸附^[5,21,93-94]。与重金属离子相结合的主要官能团有羧基、磷酸基、羟基、硫酸酯基、氨基和酰胺基,其中氮、氧、磷、硫可作为配位原子与重金属离子配位络合。络合反应在秸秆类生物炭吸附重金属机制中有重要的作用。PAN等^[95]发现4种秸秆生物炭(花生秸秆、大豆秸秆、油菜秸秆、水稻秸秆)对Cr(III)的吸附量与其含氧官能团含量成正比,并且利用红外光谱分析吸附重金属前后的生物炭表面官能团,发现吸附前后出现了不同程度的官能团位点的迁移,说明这些含氧官能团均参与了生物炭对重金属离子的吸附过程。

值得注意的是,表面络合主要是含氧官能团尤其是羧基和酚羟基^[37,82,96],会与重金属形成络合物从而参与吸附反应^[85,97]。在LU等^[85]的研究中,通过FTIR对吸附前后的污泥生物炭进行光谱分析,发现羧基和羟基官能团含量发生变化,而羰基的量几乎没有变化。SHEN等^[89]在研究中指出,硬木生物炭含有大量的羧基,会形成重金属络合物。以上研究结果说明羧基和酚羟基在与重金属的络合反应中起到了重要作用。

3.3 离子交换

离子交换反应需要在适宜的pH值条件下才能发生。酸性生物炭表面电离的质子与重金属离子发生离子交换反应^[98-99]。可以与重金属离子发生交换的电离质子包括Ca²⁺、K⁺、Mg²⁺、Si⁺、Na⁺^[61],见图1。这些无机矿物成分参与了生物炭吸附重金属的反应。LU等^[85]对Pb²⁺吸附完成后的溶液进行检测,发现有大量的Ca²⁺、K⁺、Mg²⁺、Na⁺离子被释放出来,并且释放的离子总量越高时,Pb(II)的吸附量越大。这说明在生物炭与重金属接触时,溶液

中的重金属离子会与生物炭表面电离的质子发生交换,造成电离的质子被释放到溶液中。但同时,若电离质子Ca²⁺、K⁺、Mg²⁺、Si⁺、Na⁺在溶液中存在量过大,则会对重金属的吸附造成抑制^[65],并且随离子浓度的增大,对吸附反应的抑制作用也越强。这是由于溶液中的这些离子会与重金属目标物竞争生物炭表面的离子交换位点^[100],直至达到吸附平衡。

生物炭的这些离子多存在于其表面的含氧官能团中,含氧官能团增多会增加生物炭与重金属离子发生离子交换的可能性。LIU等^[68]采用300℃水热炭化和700℃热解两种方式制备松木生物炭,将其进行表征并应用于水体中Cu(II)的吸附。结果显示,与松木原料相比,水热炭相比于原料提高了95%的含氧官能团,而热解炭降低了56%的含氧官能团。尽管水热炭的比表面积较低,但对Cu(II)的吸附能力远高于热解炭,这是因为水热炭吸附机理中离子交换反应占据主导地位。

3.4 化学沉淀

生物炭中的矿物质成分,如CO₃²⁻、PO₄³⁻、SiO₃⁴⁻,以及Cl⁻、SO₄²⁻、SO₃²⁻、OH⁻也在重金属吸附中发挥着重要作用。这些矿物质成分会与重金属离子结合形成重金属类盐沉淀物。戴静等^[61]用XRD分析了吸附重金属后的木屑、米糠、稻秆、玉米秸秆生物炭,结果显示Pb(II)和Cd(II)在生物炭表面以碳酸盐、磷酸盐、硅酸盐和亚硫酸盐形式存在。XU等^[99]将相同热解条件下制得的稻壳和牛粪生物炭应用于水体中Pb(II)、Cu(II)、Zn(II)、Cd(II)的吸附,牛粪生物炭对4种重金属吸附效果优于稻壳生物炭,分析认为,稻壳生物炭吸附重金属主要是酚羟基起作用,而牛粪生物炭吸附重金属除了羟基作用外还有大量的CO₃²⁻、PO₄³⁻参与了吸附反应。生物质原料中的CO₃²⁻、PO₄³⁻、SiO₃²⁻在与重金属Pb(II)接触时,会提供反应位点,形成Pb₅(PO₄)₃Cl^[101]、Pb₅(PO₄)₃OH和Pb₃(CO₃)₂(OH)₂^[74]、5PbO·P₂O₅·SiO₂^[85],如Pb(II)与P的共沉淀反应方程式为:6HPO₄²⁻+9Pb²⁺+6OH⁻→Pb₉(PO₄)₆+6H₂O^[102]。因此,生物炭中P、Si、Cl、S元素含量越多越有助于重金属的吸附。这也是畜禽粪便类生物炭对重金属有较高吸附量的原因之一。

值得注意的是,不可溶的P含量与吸附重金属无关,这是因为形成的P-Ca-Mg结构稳定的晶体物质(如磷钙矿(Ca,Mg)₃(PO₄)₂)不会与重金属离子发生沉淀反应^[90]。畜禽粪便生物炭的可溶性P含量在25~200℃时随着热解温度的升高而升高,原因可以归结为可溶性P质量几乎不变但生物炭

总量降低;在 200 ~ 500℃ 时随着热解温度的升高,可溶性 P 含量下降,是因为不可溶 P - Ca - Mg 晶体物质逐渐形成;在 500℃ 可以观察到磷钙矿的存在。在 200℃ 热解温度下得到的牛粪生物炭与 Pb(II) 反应后会沉积 β - $\text{Pb}_9(\text{PO}_4)_6$ 晶体;在 350℃ 热解温度下得到的牛粪生物炭与 Pb(II) 反应后除了沉积 β - $\text{Pb}_9(\text{PO}_4)_6$ 晶体外还有 $\text{Pb}_3(\text{CO}_3)_2(\text{OH})_2$ 的形成,这是因为随着热解温度的升高,生物炭碱性和方解石 (CaCO_3) 含量逐渐增强^[90]。因此,化学沉淀作用是有关组分共同作用的结果。

目前,可以通过 VisualMINTEQ 软件模拟分析化学吸附机理作用的强弱。畜禽粪便类生物炭吸附重金属,主要是重金属与畜禽粪便物料中的 CO_3^{2-} 、 PO_4^{3-} 发生反应,同时表面络合反应作用较少,化学沉淀反应占主导地位。XU 等^[27] 通过 VisualMINTEQ 模拟分析,得出牛粪生物炭对水溶液中 Pb(II) 的去除中,约 75% 以上都是化学沉淀作用,而其他吸附作用如表面络合则仅占 25%。在化学沉淀反应中约有不到 30% 的磷酸盐形成,70% 以上的碳酸盐形成,并且随着热解温度的升高,碳酸盐沉淀物生成的比重增多^[27]。倪群丽^[103] 采用 VisualMINTEQ 分析方法研究了鸡粪生物炭和猪粪生物炭对 Pb(II) 的化学吸附作用,发现猪粪生物炭对铅的吸附主要以化学沉淀为主,铅的沉淀物为 $\text{Pb}_5(\text{PO}_4)_3\text{Cl}$ 以及 Pb-碳酸盐类沉淀,如 $\text{Pb}_3(\text{CO}_3)_2(\text{OH})_2$ 、 $\text{Pb}_4(\text{SO}_4)(\text{CO}_3)_2(\text{OH})_2$, 较高热解温度制得的生物质炭更利于结晶型的 Pb-碳酸盐类化合物的形成;鸡粪生物炭对铅的吸附机理表现为表面吸附(有机络合或离子交换)和沉淀作用,沉淀物主要为硫酸铅和磷酸盐铅,磷酸盐铅的形成受溶解性磷的释放动力学过程控制。

3.5 静电吸附

静电吸附作用多发生在适宜的 pH 值条件下。当溶液 pH 值大于生物炭零电荷点 (pHpzc) 时,生物炭表面携带的负电荷能与带正电荷的重金属离子发生静电吸附作用^[95,98,104-105]。当溶液 pH 值小于生物炭 pHpzc 时,生物炭表面会电离出质子,与带负电荷的重金属离子发生静电吸附作用。因此,静电吸附作用与重金属离子电动势,在水溶液中的存在形式以及生物炭表面的电荷量密切相关,而生物炭表面负电荷量又与生物炭表面含氧官能团,包括羧基 ($-\text{COO}^-$)、羰基 ($-\text{CO}^-$) 和羟基 ($-\text{OH}$)^[106]、溶液 pH 值紧密相关^[53]。

SHEN 等^[89] 将制备的硬木生物炭用于水体中 4 种重金属的吸附,通过 Langmuir 吸附模型计算不同重金属离子吸附能力,由大到小分别为 Pb(II)、

Cu(II)、Ni(II)、Zn(II),其中对 Pb(II) 的吸附远高于其他 3 种重金属离子的原因是,Pb(II) 的电负性常数远高于其他 3 种 (Pb(II)、Cu(II)、Ni(II) 和 Zn(II) 的电负性常数分别为 2.33、1.90、1.93、0.69)^[93],因此硬木生物炭对 Pb(II) 有很强的静电吸附作用。果树枝生物炭^[21] 对 Pb(II) 和 Cr(III) 的吸附远大于 Cu(II),这是因为 Pb(II) 和 Cr(III) 有着很高的电荷密度,远高于 Cu(II),因此,Pb(II) 和 Cr(III) 与生物炭之间的库伦引力更大^[93]。

静电吸附在秸秆生物炭吸附重金属机理中占的比重小于化学沉淀。TONG 等^[63] 通过慢速热解制备得到花生秸秆、大豆秸秆和油菜秸秆生物炭,虽然油菜秸秆生物炭表面负电荷最多,但是 P 含量小于花生和大豆生物炭,造成了油菜秸秆生物炭对 Cu(II) 的吸附小于花生和大豆生物炭。

4 生物炭重金属吸附工艺条件

对于特定生物炭而言,影响生物炭重金属吸附效果的重要因素包括吸附溶液 pH 值、溶液温度、吸附剂用量、吸附平衡时间和吸附质初始浓度等。除生物炭原料、制备条件以外,吸附工艺条件对吸附效果的影响也不容忽视。

4.1 吸附溶液 pH 值

溶液理化性质 pH 值是反映水体中溶解性固体和离子变化的重要指标。溶液的 pH 值会影响生物炭表面的电荷密度和金属离子的存在形态^[4,60,70,107],进而影响生物炭对重金属离子的吸附^[66,68]。研究发现,生物炭对重金属的吸附效果通常随着 pH 值的变动而变化,当 pH 值较低时,溶液中 H^+ 和 H_3O^+ 会与阳离子竞争生物炭表面的位点,所以对于重金属离子的吸附能力减弱;随着 pH 值的逐渐增加, H^+ 量减少,增加了生物炭吸附重金属离子的位点,进而吸附效果逐渐增强。当溶液中 pH 值超过重金属离子微沉淀所需 pH 值的上限时,溶液中的大部分重金属离子会形成不溶的金属氧化物或氢氧化物颗粒,从而导致吸附无法进行。因此,溶液 pH 值过低和过高都不利于对重金属离子的吸附。不同生物炭对同种重金属的吸附随 pH 值变化存在差异^[79],并且同种生物炭对不同重金属吸附的最优 pH 值条件也不同。应在吸附应用中将溶液初始 pH 值的影响考虑在内。根据目前研究结果,大部分重金属的吸附是随着 pH 值的增加而吸附量增多,增加幅度略有不同,在 pH 值为 5 左右时,吸附量达到最大^[61,63,85]。但 Cr(III)、Cr(VI) 与其他金属不同^[51,82,108-110],QIAN 等^[110] 在 400℃ 制备了水稻

秸秆生物炭,发现在溶液 pH 值为 3.9 时对 Cr(Ⅲ)的吸附量最大。WANG 等^[51]研究了菠萝皮生物炭在 pH 值为 2~10 时对 Cr(Ⅵ)的吸附,研究得出,对 Cr(Ⅵ)的吸附量最大时 pH 值为 2,非常接近菠萝皮生物炭的等电荷点 pH 值为 1.5。当溶液 pH 值远大于等电荷点时,带负电的生物炭表面会与 Cr(Ⅵ)(CrO_4^{2-} 、 $\text{Cr}_2\text{O}_7^{2-}$)产生静电排斥作用,造成对 Cr(Ⅵ)的吸附下降。DONG 等^[82]、ZHOU 等^[37]和 TYTLAK 等^[111]也有相同的研究结果。

4.2 溶液温度

温度主要是通过影响生物炭吸附热力学过程和吸附热容等因素来进一步影响吸附效果。一般来说,吸附过程是吸热反应,可通过控制体系温度来提高吸附量^[66,107,112-114]。MOHAN 等^[10]将橡木和橡树生物炭应用于水体中重金属离子吸附,研究结果表明,溶液的温度越高生物炭对重金属吸附能力越强。张双圣等^[115]研究污泥生物炭吸附剂对 Pb(Ⅱ)的吸附效果随溶液温度的升高而增强,当溶液温度超过 30℃,吸附效果增强不明显。LIU 等^[66]发现重金属离子被吸附到生物炭表面需要足够的能量完成这一迁移过程,即生物炭吸附重金属的过程是吸热反应,因而溶液温度升高有利于吸附反应的发生。KOŁODYŃSKA 等^[60]研究结果也证明了这一点。

4.3 吸附时间

当生物炭吸附重金属时,只有保证吸附时间的充裕,才能使吸附达到平衡,充分利用生物炭吸附性能。通常生物炭重金属的吸附过程可以分为两个阶段^[116-118]:第一阶段,在生物炭与重金属接触的前几分钟,吸附率和吸附量都快速增加,这是因为吸附位点几乎完全暴露;随着吸附位点被占据,吸附率变慢随即进入吸附的第二阶段,一般持续时间较长,随后达到吸附平衡,多为 1~4 h^[85,60,118];当吸附时间超过 24 h 后,吸附量基本保持稳定。因此,在实际研究中,同时考虑时间效率和吸附效果,通常选用平衡时间作为吸附时间。

同一种材料吸附不同重金属时,吸附平衡时间也不尽相同。KOŁODYŃSKA 等^[60]在研究猪粪和牛粪对 4 种重金属(Pb、Cd、Zn、Cu)的吸附行为时,发现对 Cu 和 Zn 的吸附平衡时间是 30~60 min,对 Cd 和 Pb 的吸附平衡时间为 120~180 min。平衡时间除了与生物炭和目标重金属性质相关,还与它们的初始浓度密切相关^[119]。

4.4 吸附剂用量与重金属的初始浓度

在吸附实验中,经常通过加大生物炭的量来增加与重金属离子接触的吸附点位,进而提高对重金属的去除效果^[120]。然而,吸附剂增加过量后,由于

吸附剂本身官能团对重金属的竞争吸附能力饱和或者吸附位点达到饱和,不仅造成浪费,还影响单位质量吸附剂的吸附能力。

PELLERA 等^[118]研究了稻壳、橄榄渣、橘子皮、粪便生物炭对 Cu(Ⅱ)的吸附,4 种生物炭的用量在 2.5~25 g/L 时,随着用量的增加,吸附量增加。生物炭用量从 2.5 g/L 增加到 5 g/L 时的吸附率远大于 5 g/L 增加到 25 g/L,研究最终选择 5 g/L 的吸附剂用量。CHEN 等^[112]在研究玉米秸秆生物炭吸附重金属 Cu(Ⅱ)时也发现相同规律,优化的生物炭用量也是 5 g/L。

重金属初始浓度和吸附剂浓度对吸附量的影响有着相似的规律^[118,120-121]。AMAN 等^[121]研究了马铃薯皮生物炭对 Cu(Ⅱ)的吸附,当 Cu(Ⅱ)的初始质量浓度在 150~400 mg/L 增加时,吸附速率逐渐降低。相同的规律也被 PELLERA 等^[118]和 EL-ASHTOUKHY 等^[120]得到。因此,实验中往往需要找到同时满足吸附率和吸附量的吸附剂和重金属浓度。

5 不同目标重金属吸附能力比较分析

特定种类生物炭在吸附不同种重金属时,对每种重金属的吸附能力也有差异^[60]。INYANG 等^[5]采用消解甘蔗渣(DWSBC)通过热解方式制备生物炭应用于水体中 Pb(Ⅱ)、Cu(Ⅱ)、Ni(Ⅱ)和 Cd(Ⅱ)的吸附,DWSBC 对 4 种重金属吸附能力由小到大依次为 Pb、Cu、Cd、Ni。UCHIMIYA 等^[98]分别在 350℃ 和 700℃ 条件下慢速热解制备了鸡粪生物炭,结果显示,对 Pb(Ⅱ)、Cu(Ⅱ)、Ni(Ⅱ)和 Cd(Ⅱ)4 种重金属吸附能力由小到大依次为 Pb、Cu、Cd、Ni。RAFATULLAH 等^[122]研究发现,柳桉木生物炭对不同重金属的吸附能力由小到大为 Pb、Ni、Cu、Cr。PARK 等^[40]制备的芝麻秸秆生物炭对 5 种重金属的最大吸附量由小到大依次为 Pb(102 mg/g)、Cd(86 mg/g)、Cr(65 mg/g)、Cu(55 mg/g)、Zn(34 mg/g)。MOHAN 等^[119]制备了橡木树皮生物炭,其单位表面积上对水中 Pb(Ⅱ)、Cd(Ⅱ)的吸附量分别为 0.515 7、0.213 mg/m²。根据目前已有研究表明,相较于其他重金属,生物炭对 Pb(Ⅱ)表现出较高的吸附能力。

6 总结与展望

生物炭作为一种新型的吸附材料,具有绿色环保、价格低廉、来源广泛、吸附性能强等特点,使得其在重金属吸附方面表现出良好的潜力和前景。目前,虽然针对生物炭重金属的吸附研究取得了不少进展,但未来的研究工作还应侧重以下几方面。

(1) 在生物炭特性及其吸附产物方面, 针对最优吸附能力的物料选择和工艺参数缺乏研究^[123], 因此, 应开展针对重金属吸附的生物炭原料特性及吸附产物的多维、微纳尺度表征方法研究, 以期生物炭重金属吸附剂的定向生产奠定基础。

(2) 在生物炭的吸附特性方面, 由于生物炭的制备原料和制备条件不同, 吸附也多在不同初始 pH 值、温度、吸附剂浓度、吸附质浓度等工艺条件下展开, 所揭示的吸附效果均是在特定制备条件和吸附条件下, 可比性较弱^[124]。应将生物炭特性结合其吸附效果进行相关性分析。

(3) 在吸附机理研究方面, 已有研究工作尚处于定性描述阶段, 对生物炭吸附重金属过程中, 各个吸附机制对吸附过程的贡献以及生物炭表面官能团对吸附贡献的比重还不系统、详细。研究仅停留在实验室的模拟与分析上。应采用多维、微纳尺度表征手段分析吸附机制和吸附过程。

(4) 已有研究多局限于生物炭对单一重金属溶液体系的吸附, 而实际污染水体多为复合重金属污染^[123]。当多种重金属同时被吸附时, 是否产生相互作用, 以及在重金属复合污染物体系中, 吸附复杂机理研究应是未来的一个研究重点。

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